



Effects of dust aerosols on tropospheric chemistry during a typical pre-monsoon season dust storm in northern India

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Abstract. This study examines the effect of a typical pre-monsoon season dust storm on tropospheric chemistry through a case study in northern India. Dust can alter photolysis rates by scattering and absorbing solar radiation and provide surface area for heterogeneous reactions. We use the Weather Research and Forecasting model coupled with Chemistry (WRF-Chem) to simulate the dust storm that occurred during 17–22 April 2010 and investigate the contribution of different processes on mixing ratios of several key trace gases including ozone, nitrogen oxides, hydrogen oxides, methanol, acetic acid and formaldehyde. We revised the Fast Troposphere Ultraviolet Visible (F-TUV) photolysis scheme to include effects of dust aerosols on photolysis rates in a manner consistent with the calculations of aerosol optical properties for feedback to the meteorology radiation schemes. In addition, we added 12 heterogeneous reactions on the dust surface, for which 6 reactions have relative-humidity-dependent reactive uptake coefficients (γ). The inclusion of these processes in WRF-Chem is found to reduce the difference between observed and modeled O_3 from 16 ± 9 to 2 ± 8 ppbv and that in NO_y from 2129 ± 1425 to 372 ± 1225 pptv compared to measurements at the high-altitude site Nainital in the central Himalayas, and reduce biases by up to 30 % in tropospheric column NO_2 compared to OMI retrievals. The simulated dust storm acted as a sink for all the trace gases examined here and significantly perturbed their spatial and vertical distributions. The reductions in these gases are estimated as 5–100 %, and more than 80 %

of this reduction was due to heterogeneous chemistry. The RH dependence of γ is also found to have substantial impact on the distribution of trace gases, with changes of up to 20–25 % in O_3 and HO_2 , 50 % in H_2O_2 and 100 % in HNO_3 . A set of sensitivity analyses revealed that dust aging could change H_2O_2 and CH_3COOH levels by up to 50 % but has a relatively small impact on other gases.

1 Introduction

Dust aerosols have gained considerable attention in the recent years not only because they constitute a major fraction of the particulate matter in the troposphere but also because they have important implications for air quality, visibility, the earth's radiation budget (e.g., Haywood and Boucher, 2000; Seinfeld et al., 2004), biogeochemistry (e.g., Jickells et al., 2005), hydrological cycles (e.g., Miller et al., 2004; Zhao et al., 2011), and atmospheric chemistry (e.g., Dentener et al., 1996; Wang et al., 2012). The significance of dust aerosols for atmospheric chemistry has been manifested through several experimental (e.g., Goodman et al., 2000; Underwood et al., 2001; Li et al., 2006; Preszler Prince et al., 2007; Wagner et al., 2008; Cwiertny et al., 2008; Pradhan et al., 2010; Crowley et al., 2010; Chen et al., 2011; Bedjanian et al., 2013a, b) and modeling studies (e.g., Zhang et al., 1994; Dentener et al., 1996; Zhang and Carmichael, 1999; Tang et al., 2004; Martin et al., 2003; Bauer et al., 2004; Tie et al.,

2005; Hodzic et al., 2006; Zhu et al., 2010; Wang et al., 2012) during the past 2 decades.

Dust aerosols can influence atmospheric chemistry by affecting the photolysis rate coefficients through interaction with incoming solar radiation and by providing surface area for heterogeneous chemistry and deposition of different trace gases. It has been suggested that mineral dust aerosols are responsible for 5–20 % reduction in photolysis rate coefficients of ozone (O₃) and nitrogen dioxide (NO₂) throughout the Northern Hemisphere (e.g., Martin et al., 2003; Tie et al., 2005; Ying et al., 2011). These changes in photolysis rate coefficients then decrease the annual mean global concentration of hydroxyl radical (OH) by 9 %, which in turn leads to increases in the concentrations of several volatile organic compounds (Martin et al., 2003). Heterogeneous reactions on dust surfaces generally reduce the concentration of key atmospheric trace gases such as O₃, nitrogen oxides, sulfur oxides and hydrogen oxides, but the amount of reported reduction in these gases varies widely, indicating that heterogeneous chemistry on dust surfaces is still not well understood. For example, the decreases in surface O₃ are reported as 5–40 % by different studies (Zhang et al., 1994; Dentener et al., 1996; Zhang and Carmichael, 1999; Tang et al., 2004; Bauer et al., 2004; Zhu et al., 2010; Wang et al., 2012). Likewise, the decreases for sulfur dioxide (SO₂), nitrogen oxides and hydrogen oxides are reported as 10–50, 16–100 and 11–59 % respectively.

Even though this study is focused on the impact of heterogeneous processes on dust surfaces on the tropospheric chemistry in northern India, it is worth mentioning other implications of heterogeneous chemistry here. The heterogeneous uptake of trace gases can potentially affect the physicochemical properties of dust aerosols and enhance their ability to act as cloud condensation nuclei (CCN). Dust particles are generally hydrophobic when they are emitted but become hygroscopic as they travel in the atmosphere and become coated with nitrate, sulfate and organics (e.g., Levin et al., 1996; Kelly et al., 2007; Hatch et al., 2008). The coating process will also modify the distributions of dust as well as sulfate and nitrate aerosols, which in turn will increase the scattering of solar radiation by aerosols. The increase in size of dust particles due to coating will increase the probability of their removal from the atmosphere, and such coated particles will experience less long-range transport.

Box and regional modeling studies have focused mostly on heterogeneous chemistry on East Asian and Saharan dust aerosols, whereas there have been no such studies over the northern Indian region where dust storms occur frequently during the pre-monsoon (March, April, May; MAM) season (e.g., Prasad and Singh, 2007; Hegde et al., 2007). Global modeling studies have suggested that heterogeneous chemistry on dust aerosols can reduce surface O₃ in northern India by 4–10 % (e.g., Dentener et al., 1996; Bauer et al., 2004). However, there has been a considerable improvement in our understanding of heterogeneous chemistry on dust aerosols

since these global modeling studies were conducted. A major advancement has been the demonstration of relative humidity (RH) dependence of the reactive uptake coefficient for several species such as O₃ (Cwierny et al., 2008), HNO₃ (Liu et al., 2008), OH (Bedjanian et al., 2013a), HO₂ (Bedjanian et al., 2013b), SO₂ (Preszler Prince et al., 2007) and H₂O₂ (Pradhan et al., 2010). The renoxification process (Eq. 1), which was ignored in those global modeling studies, is also included here. In this process, the uptake of gas-phase HNO₃ by dust particles is followed by release of gas-phase NO_x (both NO and NO₂) in the presence of broadband radiation (e.g., Chen et al., 2011).



where (g) and (p) represent the species in gas and particle phase respectively. Therefore, it is essential to reassess the importance of dust aerosols for tropospheric chemistry in northern India by taking into account these recent advancements.

In light of the above conditions, this manuscript examines the effects of dust aerosols on the distribution of many key trace gases including O₃, nitrogen oxides, hydrogen oxides, methanol, acetic acid and formaldehyde by incorporating the updated information on heterogeneous reactive uptake of trace gases in MOZCART chemical mechanism of Weather Research and Forecasting model coupled with Chemistry (WRF-Chem). We also revise the Fast Troposphere Ultraviolet Visible (F-TUV) scheme to include effects of dust aerosols on photolysis rates. This extended configuration of WRF-Chem is then used to simulate the impact of a typical pre-monsoon season dust storm on the regional tropospheric chemistry in northern India. This dust storm occurred during 17–22 April 2010 in northern India. A detailed analysis of the evolution of this dust storm, dust emissions and their effect on local to regional-scale aerosol optical properties and radiation budget is presented in a companion paper (Kumar et al., 2014).

We begin by describing the WRF-Chem configuration and implementation of effect of dust aerosols on photolysis rates and heterogeneous chemistry on dust surfaces. We next briefly describe the observations used to evaluate the WRF-Chem trace gas results. In the Results section, we show the importance of dust effects on tropospheric gas chemistry by comparing the model results with observations. Sensitivity simulations are used to discuss the contributions of different processes and evaluate the uncertainties in the representation of heterogeneous chemistry. We conclude that heterogeneous chemistry mostly, partly dust-modified photolysis rates, improves the agreement between model results and observations, and that these processes lead to a reduction in O₃, H₂O₂, and HNO₃.

2 The WRF-Chem model

The Weather Research and Forecasting Model version 3.4.1 (Skamarock et al., 2008) coupled with Chemistry (Grell et al., 2005; Fast et al., 2006) is used here to simulate the distribution of trace species. The simulation domain has $120 \times 90 \times 51$ grid points in (x, y, z) direction with a grid spacing of $30 \times 30 \text{ km}^2$ extending in the vertical up to 10 hPa. The static geographical fields, such as terrain height, vegetation fraction, soil properties and land use, etc., are interpolated from 10 min ($\sim 19 \text{ km}$) United States Geological Survey (USGS) data to the model domain by using the WRF preprocessing system (WPS). The initial and lateral boundary conditions for the meteorological fields are obtained from the National Center for Environmental Predictions (NCEP) Final Analysis (FNL) fields available every 6 h at a spatial resolution of $1^\circ \times 1^\circ$. The resolved scale cloud physics is represented by the Thompson microphysics scheme (Thompson et al., 2004), and subgrid-scale effects of convective and shallow clouds are parameterized according to the Kain–Fritsch convective scheme (Kain, 2004). The short- and long-wave radiative processes in the atmosphere are represented by the Rapid Radiative Transfer Model (RRTM) (Mlawer et al., 1997). For surface processes, the model setup uses the NOAA Land Surface model (Chen and Dudhia, 2001) and MM5 similarity scheme (Beljaars, 1994). The vertical subgrid-scale fluxes due to eddy transport in the planetary boundary layer (PBL) and the free troposphere are parameterized according to the Yonsei University (YSU) boundary layer scheme (Hong et al., 2006). Four-dimensional data assimilation (FDDA) is applied to limit model errors in the simulated meteorological fields (Lo et al., 2008).

Gas-phase chemistry is represented by the Model for Ozone and Related chemical Tracers, version 4 (MOZART-4, Emmons et al., 2010), and aerosol processes by the GOCART bulk aerosol scheme (Chin et al., 2002; Pfister et al., 2011). The GOCART model simulates five major tropospheric aerosol types including sulfate, organic carbon, black carbon, dust and sea salt, assuming externally mixed aerosols. The GOCART model does not have an aerosol thermodynamics module. The emissions of sea salt (four size bins) and dust aerosols (five size bins) are calculated online within the model. The dust emission scheme is based on Ginoux et al. (2001) and calculates size-resolved dust emissions online in five size bins ranging from 0.73 to $8.0 \mu\text{m}$ (effective radius) using the following equation:

$$F_p = \begin{cases} C S s_p u_{10\text{m}}^2 (u_{10\text{m}} - u_t) & \text{if } u_{10\text{m}} > u_t \\ 0 & \text{otherwise} \end{cases}, \quad (2a)$$

where F_p ($\text{kg m}^{-2} \text{s}^{-1}$) represents the emission flux for size bin p , C is an empirical proportionality constant ($\text{kg m}^{-5} \text{s}^2$), S is the source function representing the fraction of alluvium available for wind erosion, s_p is the fraction of each size class of dust in the emission, $u_{10\text{m}}$ (m s^{-1}) is the horizontal wind

speed at 10 m above the surface and u_t is the threshold velocity (m s^{-1}) below which dust emission does not occur and is a function of particle size and density, air density and surface moisture. In principle, S should be based on observations of alluvium in the model domain, but such observations do not exist. Therefore, S is obtained by comparing elevation of each grid cell with its surrounding hydrological basin spread over an area of $10^\circ \times 10^\circ$. Thus, S does not provide the exact amount of alluvium present, but it provides information about the most probable locations of sediments. The map of S is then compared with vegetation map derived from Advanced Very High Resolution Radiometer (AVHRR) data (DeFries and Townshend, 1994), and only the bare soil surfaces are considered as possible dust sources. In WRF-Chem, the source function is provided as an input static geographical field through the WRF preprocessing system (WPS). The threshold wind velocity u_t is estimated as

$$u_t = \begin{cases} A \sqrt{\frac{(\rho_p - \rho_a) g \Phi_p}{\rho_a}} (1.2 + 0.2 \log_{10} w), & \text{if } w < 0.5, \\ \infty & \text{otherwise} \end{cases} \quad (2b)$$

where $A = 6.5$ is a dimensionless parameter, w is the surface wetness (0.001–1), Φ_p is the particle diameter, g is the acceleration of gravity, and ρ_a and ρ_p are the air and particle density respectively. More information about estimation of S and u_t can be obtained from Ginoux et al. (2001).

The tuning factor C does not have any physical meaning. Its default value was proposed initially as $1 \times 10^{-9} \text{ kg m}^{-5} \text{ s}^2$ (Ginoux et al., 2001), but we have set it to $2.2 \times 10^{-8} \text{ kg m}^{-5} \text{ s}^2$ as the latter value led to a good agreement between model and Aerosol Robotic Network (AERONET) measured aerosol optical depth and Angström exponent at seven sites in the model domain (Kumar et al., 2014). A factor of magnitude difference in the value of C results partly from the use of analysis nudging (FDDA) in this study as nudging leads to lower wind speeds. A sensitivity experiment showed that similar dust emissions could be obtained with a C value of $9 \times 10^{-9} \text{ kg m}^{-5} \text{ s}^2$ without applying analysis nudging. However, we chose to use analysis nudging as the model configuration with analysis nudging is found to better reproduce the variations in observed aerosol optical properties as compared to one without analysis nudging (not shown).

Anthropogenic emissions of CO, NO_x, SO₂, NH₃, OC and BC and non-methane volatile organic compounds (NMVOCs) are taken from the MACCity emission inventory (Granier et al., 2011) and emissions for PM_{2.5} and PM₁₀ are taken from the Intercontinental Chemical Transport Experiment – Phase B (INTEX-B) inventory (Zhang et al., 2009). Daily varying emissions of trace species from biomass burning are taken from the Fire Inventory from NCAR version 1 (FINN v1) (Wiedinmyer et al., 2011) and distributed vertically in the model following the online plume-rise module (Freitas et al., 2007). Note that FINN v1 accounts only for

open biomass burning, and the residential biomass burning is included in anthropogenic emissions. Biogenic emissions of trace species from terrestrial ecosystems are calculated on-line using the Model of Emissions of Gases and Aerosols from Nature (MEGAN) version 2.04 (Guenther et al., 2006). The aerosols are allowed to provide feedback to the radiation scheme in the simulations but through direct effects only. The simulations start on 10 April 2010 at 00:00 GMT and end on 25 April 2010 at 18:00 GMT. The first 3 days of the model output are removed from the analysis to allow the model to spin up.

2.1 Effect of aerosols on photolysis rates in WRF-Chem

In this study, we use the Fast-Troposphere Ultraviolet Visible (F-TUV) scheme to calculate photolysis rates for the MOZ-CART chemical mechanism. The F-TUV scheme is a simplified version of the National Center for Atmospheric Research's (NCAR) TUV model (Madronich and Weller, 1990) and is designed to reduce the computational costs associated with TUV (Tie et al., 2005). The F-TUV model utilizes the same physical processes as the TUV model except that the number of wavelengths in the TUV spectra (121–750 nm) was reduced from 140 to 17, making it eight times faster than the TUV model. The differences in the calculated photolysis rate coefficients between the TUV and F-TUV model are reported to be less than 5 % (Tie et al., 2005).

To include the effect of aerosols on photolysis rates, the F-TUV photolysis scheme in WRF-Chem calculates optical properties (optical depth, single scattering albedo and asymmetry parameter) for black carbon, organic carbon, sulfate and sea-salt aerosols, and passes them to a two-stream radiative transfer module where they interact with radiation to affect photolysis rate coefficients. These optical properties are also calculated in the aerosol optical driver of WRF-Chem and are used for feedback of aerosols on the meteorology radiation schemes. Here, we make use of the optical driver calculated aerosol optical properties in F-TUV photolysis scheme to be consistent between the effects of aerosols on radiation and photolysis rate coefficients. This coupling automatically accounts for the effect of dust aerosols on photolysis rates in WRF-Chem, which was missing in the original F-TUV scheme. Further details regarding the coupling of the optical driver to the F-TUV photolysis scheme can be found in the supplementary material.

2.2 Heterogeneous chemistry on dust surface in WRF-Chem

This study also extends the ability of the MOZCART chemical mechanism of WRF-Chem to simulate heterogeneous chemistry on the surface of dust particles by including 12 heterogeneous reactions listed in Table 1. The uptake of these gases on dust particles is assumed to be irreversible (Zhang and Carmichael, 1999) and does not add to the aerosol mass

in the model because the GOCART model does not have a thermodynamics module. However, recent laboratory experiments have shown that the uptake of HNO₃ (Chen et al., 2011) is associated with release of gas-phase NO_x, and uptake of OH (Bedjanian et al., 2013a) and HO₂ (Bedjanian et al., 2013b) are associated with the release of gas-phase H₂O₂. The production of these gas-phase species from heterogeneous chemistry is taken into account in this study. The production of NO_x from HNO₃ uptake is observed only in the presence of broadband irradiation with a yield of about 50 % (Chen et al., 2011), and thus the reaction of HNO₃ is set to yield 0.5 NO₂ during daytime in this study. This yield of 0.5 results from the assumption that all HNO₃ molecules reacting with dust particles adsorb to the surface, but HNO₃ has been observed to be lost irreversibly when reacting with CaCO₃ and adsorb to the surface when reacting with other minerals (e.g., Al₂O₃, Fe₂O₃, and TiO₂) (Grassian, 2002). For Indian dust particles, the abundance of CaCO₃ is estimated to be about 20 % (Peterson, 1968). If we assume that HNO₃ will react with the remaining 80 % of the minerals adsorbing to the surface, the yield should be 0.4. However, our assumption of a yield of 0.5 should not affect the model results significantly here as it is shown later (Sect. 4.3) that the renoxification process leads only to a small change in NO_x values. The yields for H₂O₂ from reactions of OH and HO₂ with dust particles are reported to be 5 % (Bedjanian et al., 2013a) and 10 % (Bedjanian et al., 2013b) respectively and are set accordingly. These numbers differ from previous studies, which have assumed a 100 % conversion of HO₂ and 0 % conversion of OH into H₂O₂ (e.g., Zhu et al., 2010; Wang et al., 2012).

The pseudo-first-order reaction rate coefficient (s⁻¹) for the loss of a gas-phase species *g* due to heterogeneous uptake by dust particles is calculated following Heikes and Thompson (1983) as

$$k_g = \sum_{i=1}^5 \frac{4\pi r_i D_g V N_i}{1 + Kn[\chi + 4(1 - \gamma)/3\gamma]}, \quad (3)$$

where *i* = 1, 5 represents five dust size bins used by the GOCART model, *r_i* and *N_i* represents the effective radius (cm) and number density (particles cm⁻³) of particles in size bin *i*, *V* is the ventilation coefficient and taken as unity, and *D_g* represents the gas-phase molecular diffusion coefficient of gas molecule *g* (cm² s⁻¹) and is calculated following Jacobson (2005) as

$$D_g = \frac{5}{16\rho_a A d_g^2} \sqrt{\left(\frac{m_a + m_g}{m_g}\right) \frac{RT m_a}{2\pi}}, \quad (4)$$

where ρ_a is atmospheric mass density (g cm⁻³), *A* is the Avogadro number (6.022 × 10²³ molecules mol⁻¹), *d_g* is the collision diameter (cm) of gas molecule *g*, *R* is the universal gas constant (8.31451 × 10⁷ g cm² s⁻² mol⁻¹ K⁻¹), *T* is the absolute temperature (K), and *m_a* (g mol⁻¹) and *m_g* (g mol⁻¹)

Table 1. The heterogeneous reactions and dry reactive uptake coefficients (γ_{dry}) used in this study are listed. γ_{low} and γ_{high} represent lower and upper bounds of reactive uptake coefficient reported in the literature. The rightmost column shows the references based on which relative humidity dependence of γ_{dry} is specified.

Reaction	γ_{low}	γ_{dry}	γ_{high}	RH dependence
$\text{O}_3 + \text{Dust} \rightarrow \text{P}$	1×10^{-6}	2.7×10^{-5}	1×10^{-4}	Cwiertny et al. (2008)
$\text{HNO}_3 + \text{Dust} \rightarrow 0.5 \text{NO}_x + \text{P}$	1×10^{-5}	2.0×10^{-3}	0.2	Liu et al. (2008)
$\text{NO}_2 + \text{Dust} \rightarrow \text{P}$	4×10^{-10}	2.1×10^{-6}	2×10^{-4}	–
$\text{NO}_3 + \text{Dust} \rightarrow \text{P}$	0.01	0.1	0.23	–
$\text{N}_2\text{O}_5 + \text{Dust} \rightarrow \text{P}$	0.01	0.03	0.1	–
$\text{OH} + \text{Dust} \rightarrow 0.05 \text{H}_2\text{O}_2 + \text{P}$	0.004	0.18	1	Bedjanian et al. (2013a)
$\text{HO}_2 + \text{Dust} \rightarrow 0.1 \text{H}_2\text{O}_2 + \text{P}$	0.01	6.4×10^{-2}	1	Bedjanian et al. (2013b)
$\text{H}_2\text{O}_2 + \text{Dust} \rightarrow \text{P}$	8×10^{-4}	2×10^{-3}	0.18	Pradhan et al. (2010)
$\text{SO}_2 + \text{Dust} \rightarrow \text{P}$	5×10^{-7}	3.0×10^{-5}	2.6×10^{-4}	Preszler Prince et al. (2007)
$\text{CH}_3\text{COOH} + \text{Dust} \rightarrow \text{P}$	2.4×10^{-4}	1×10^{-3}	2×10^{-3}	–
$\text{CH}_3\text{OH} + \text{Dust} \rightarrow \text{P}$	4×10^{-6}	1×10^{-5}	1.9×10^{-4}	–
$\text{CH}_2\text{O} + \text{Dust} \rightarrow \text{P}$	2.6×10^{-7}	1×10^{-5}	1.1×10^{-4}	–

are the molecular weights of air and gas respectively. In Eq. (3), Kn is the dimensionless Knudsen number defined as the ratio of the effective mean free path of a gas molecule in air (λ) to the effective particle radius r_i . χ represents a correction factor for anisotropic movement and is calculated as follows:

$$\chi = \frac{\left(\frac{4}{3}Kn + 0.71\right)}{Kn + 1} \quad (5)$$

WRF-Chem model simulates the mass mixing ratio (M_i in g g^{-1}) of dust particles, and the use of Eq. (3) requires the conversion of mass mixing ratios to the number concentration. This is performed using the following equation:

$$N_i = \frac{M_i \rho_a}{\left(\frac{4}{3}\pi r_i^3\right) \rho_p} \quad (6)$$

where ρ_p (g cm^{-3}) represents the mass density of the dust particles, which are taken as 2.5 and 2.65 g cm^{-3} to represent clay and silt in GOCART (Ginoux et al., 2001).

The reaction uptake coefficient (γ) is the most important parameter in the calculation of k_g , but the uncertainties in γ are very large and have been reported as more than 3 orders of magnitudes for certain species such as O_3 and HNO_3 (Goodman et al., 2000; Underwood et al., 2001; Michel et al., 2002). Such large uncertainties make the choice of γ very difficult. Since this study is focusing on Thar Desert dust aerosols for which information on heterogeneous chemistry kinetics is not available, we use the best guess values reported for East Asian dust aerosols (Zhu et al., 2010) for the γ values of dry dust particles except for OH and HO_2 . The γ values for OH and HO_2 are taken from Bedjanian et al. (2013a, b). The applied γ values for dry dust particles are shown in Table 1 and vary from 2.1×10^{-6} for NO_2 to 0.18 for OH. Lower and upper bounds of γ reported in

the literature are also shown in Table 1. Further information about available measurements of γ for different species can be found in Zhu et al. (2010).

Many laboratory studies have also demonstrated the dependence of γ on relative humidity (RH), but such a RH dependence has generally been ignored in previous modeling studies (e.g., Dentener et al., 1996; Zhu et al., 2010; Wang et al., 2012). In this study, we include the RH dependence of γ for O_3 (Cwiertny et al., 2008), HNO_3 (Liu et al., 2008), OH (Bedjanian et al., 2013a), HO_2 (Bedjanian et al., 2013b), H_2O_2 (Pradhan et al., 2010), and SO_2 (Preszler Prince et al., 2007). The variations of γ with relative humidity for different trace gases are shown in Fig. 1. The value of γ increases with RH for HNO_3 , SO_2 and H_2O_2 while it decreases with RH for O_3 , OH and HO_2 . The rate coefficients, i.e., k_g values, are estimated to be of the order of 10^{-3} – 10^{-5} s^{-1} for OH, HO_2 , H_2O_2 , HNO_3 , NO_3 , N_2O_5 and CH_3COOH , and of the order of 10^{-7} – 10^{-8} s^{-1} for O_3 , NO_2 , SO_2 , CH_3OH and CH_2O .

The aging of dust particles through heterogeneous uptake of acidic gases and organic compounds is another important process that can influence the uptake of trace gases by dust particles. However, the dependence of k_g on aging of mineral dust aerosols is complex and not well understood. For example, the reactive uptake of O_3 (Usher et al., 2003) on HNO_3 and SO_2 processed dust particles is reported to increase or decrease depending on the mineralogy of the particle, coverage of the coatings, as well as ambient RH. Due to the complexity of processes involved and lack of information on all 12 heterogeneous reactions, the effect of atmospheric processing of dust particles is not included in this study except for a set of sensitivity simulations conducted to highlight the importance of this aging process. The treatment of dust aging for those sensitivity simulations is presented below.

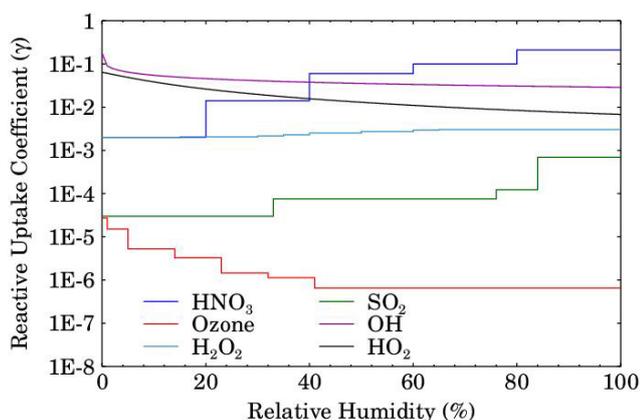


Figure 1. The relative humidity dependence of reactive uptake coefficient (γ) for O₃, HNO₃, OH, HO₂, H₂O₂ and SO₂ used in this study.

2.3 Atmospheric aging/processing of dust particles

Laboratory experiments have shown that reactive uptake of O₃ decreases by about 70 % on dust particles processed with HNO₃ (leading to nitrate coating) while it increases by about 33 % on dust particles processed with SO₂ as O₃ reacts with surface-bound sulfites/bisulfites to form sulfate (Usher et al., 2003). We have made an attempt here to simulate these changes in dust reactivity and quantify their impact on surface O₃. The simulation of these changes requires partitioning of fresh and aged dust particles and is done using the following procedure: in addition to total dust, two artificial dust tracers called “fresh_dust” and “aged_dust” are introduced into the model to keep track of fresh and aged dust particles. These tracers are included for all five size bins, and thus we have a total of 10 tracers in the model. All these tracers undergo the same transport and deposition processes as the original dust tracers.

The emissions of fresh dust particles are set equal to the emissions of total dust while the emissions of aged dust particles are set to zero. The initial and boundary conditions for fresh dust particles are set equal to those for total dust while those for aged dust particles are set to zero because the MOZART-4 output used for providing initial and boundary conditions does not include such a classification. The assumption that all dust particles entering into the model domain are fresh may introduce some uncertainty in the model results. However, such an uncertainty is anticipated to have a small contribution in the case presented here, as two model runs with and without including the regional dust emissions showed that most of the dust loading over the model domain came from emissions within the model domain during 17–22 April 2010.

The number concentration (particles cm⁻³) of aged ($N_{a,i}$, $i = 1, 5$ represent five size bins) and fresh ($N_{f,i}$) dust particles is updated every time step as follows for each gas g

considered: first, the number of molecules of gas g needed to completely coat one dust particle of a given size with a monolayer ($n_{ml,g,i}$) is calculated as the ratio of surface area of the dust particle to the surface area occupied by a gas molecule. Second, the total number of gas molecules of g that can potentially adsorb to dust particles ($n_{pot,g,i}$) is calculated by multiplying k_g (s⁻¹) estimated in Eq. (3) with gas concentration (molecules cm⁻³) and time step (180 s in this case). The ratio $n_{pot,g,i}/n_{ml,g,i}$ then provides the number concentration increment $\Delta N_{a,g,i}$ of dust particles that could have aged (i.e., been completely coated with a monolayer of gas) during this time step. The sum of the increments $\Delta N_{a,g,i}$ due to all gases considered in coating gives the total increment in aged particles $\Delta N_{a,i}$. $\Delta N_{a,i}$ is then subtracted from the number concentration of fresh dust particles $N_{f,i}$ and added accordingly to the number concentration of aged particles $N_{a,i}$.

The reactions of HNO₃, NO₂, NO₃ and N₂O₅ with dust particles are assumed to coat dust particles with nitrate and that of SO₂ is assumed to coat them with sulfate. All gases are given equal probability to react with dust particles, and in case the number concentration of fresh dust particles is limiting, then $\Delta N_{a,i} > N_{f,i}$, which would lead to negative number concentrations of $N_{f,i}$. To overcome this problem, $N_{a,i}$ and $N_{f,i}$ are set to $N_{f,i}$ and zero respectively. This approach leads to mass conservation of tracers, and the sum of fresh and aged dust particle concentrations is always equal to the total dust number concentration. Both fresh and coated dust particles then react separately with trace gases.

2.4 Simulations conducted

Eleven simulations were conducted (Table 2) to examine the impact of the dust storm on tropospheric chemistry. No_Dust serves as a base simulation in which heterogeneous chemistry on dust surfaces and the effect of dust on photolysis rates are not included. Dust_J simulation includes the effect of dust aerosols on photolysis rates, while Dust_JH simulation is same as Dust_J but with addition of heterogeneous chemistry on dust surface taking into account the RH dependence of γ and renoxification process. Dust_JH_NoRH simulation is same as Dust_JH except that it does not include the RH dependence of γ . In Dust_JH_NoReNO_x, the release of gas-phase NO₂ associated with uptake of HNO₃ is excluded in order to assess the importance of the renoxification process. Dust_JH_LoG and Dust_JH_HiG simulations are conducted with lower and upper bounds of γ reported in the literature in order to provide an estimate of the uncertainty in heterogeneous chemistry induced changes in tropospheric chemistry due to the uncertainty in γ . Dust_JH_NO₃ and Dust_JH_SO₄ simulations are designed to examine the influences of dust coated with nitrate and sulfate separately on the uptake of O₃. In these simulations, the uptake coefficient γ for aged dust particles is reduced by 70 % in Dust_JH_NO₃ and is increased by 33 % in Dust_JH_SO₄ (Usher et al., 2003). Dust particles are coated with both sulfate and nitrate

in Dust_JH_NO₃_SO₄ to examine the combined effect of nitrate and sulfate coating on O₃ uptake. In this simulation, three reactions of O₃ are included: with fresh dust particles, dust particles coated with nitrate and dust particles coated with sulfate. The original γ value is used for reaction with fresh dust particles, while the original γ value is reduced by 70 % for reaction with dust particles coated with nitrate and is increased by 33 % for reaction with dust particles coated with sulfate. Dust_JH_Sat is a hypothetical simulation where we assume that the presence of a nitrate or sulfate monolayer on the dust particle would saturate the dust particles and deactivate them for further catalytic uptake of O₃, OH, HO₂, H₂O₂, CH₃COOH, CH₃OH and HCHO, but the coating is assumed to have no effect on the uptake of HNO₃, NO₂, NO₃, N₂O₅ and SO₂. This comparison of Dust_JH_Sat with Dust_JH and Dust_J would provide bounds of the heterogeneous chemistry induced changes in tropospheric chemistry.

3 Observation data sets

3.1 Surface observations

This study uses surface O₃ and NO_y (sum of nitrogen oxides) observations made at the high-altitude site Nainital (79.45° E, 29.36° N, 1958 m a.m.s.l.) located in the central Himalayas. The observation site is bounded by high-altitude (2–5 km) mountains in the north and east directions and opens to the Indo-Gangetic Plain region in the south and west directions. There are no major anthropogenic sources near Nainital, and thus the observations at this site are envisaged to be representative of northern India (Kumar et al., 2010). Further details regarding the orography, vegetation cover, meteorological and chemical characteristics of Nainital can be found elsewhere (e.g., Sagar et al., 2004; Pant et al., 2006; Kumar et al., 2010; Sarangi et al., 2014). Ozone measurements are made using a standard ultraviolet absorption based instrument, and NO_y measurements are made with a chemiluminescence-based instrument. More details about the measurement setup, operating principle, accuracy, detection limits and calibration procedure are discussed elsewhere (e.g., Kumar et al., 2010; Sarangi et al., 2014). The model results are also compared with Aerosol Robotic Network (AERONET) measurements at seven sites in the model domain. Further details of AERONET and these observations sites can be found in Kumar et al. (2014).

3.2 Ozone Monitoring Instrument (OMI)

The Ozone Monitoring Instrument (OMI), aboard NASA's Earth Observing System (EOS) Aura satellite, measures the radiation backscattered by the earth's atmosphere and surface over the 0.27–0.5 μm wavelength range with a spatial resolution of about 13 km \times 24 km at nadir in normal operational mode. These measured radiances are used for daily global retrievals of several trace species, such as O₃, NO₂,

BrO, SO₂, CH₂O and aerosols. However, we do not use OMI O₃ and SO₂ retrievals because of their low sensitivity in the lower troposphere, which is the region of interest in this study. We find that OMI SO₂ retrievals are very noisy as average OMI-retrieved SO₂ planetary boundary layer (PBL) column amount values (0.5–1.0 DU) are smaller than the reported standard deviation of the noise (1.5 DU). Here, we use Level-2 tropospheric column NO₂ data sets made available by KNMI (Royal Netherlands Meteorological Institute) as they provide access to the averaging kernel and a priori profiles, which are needed to make a proper comparison between model profiles and satellite retrievals (e.g., Emmons et al., 2004). More details on the algorithm used to retrieve the tropospheric column NO₂ abundances at KNMI are described by Bucsela et al. (2006). These OMI NO₂ retrievals were found to correlate well with aircraft measurements made during the INTEX-B campaign (Boersma et al., 2008) and MAX-DOAS ground-based measurements (Kramer et al., 2008) but are also suggested to be biased positively by about 0–30 %, irrespective of season (e.g., Boersma et al., 2009a; Zhou et al., 2009).

To compare WRF-Chem results with OMI, the best quality OMI retrievals are used by reducing influence of clouds on OMI retrievals through selection of pixels with cloud fraction less than 0.3 and removing unreliable retrievals associated with a tropospheric column flag of greater than 0 (Boersma et al., 2009b). The nighttime pixels from OMI are also excluded for the comparison. These best quality retrievals are co-located in space and time with model output. The co-located WRF-Chem profiles are then transformed using the averaging kernel and a priori profiles used in the satellite retrievals to obtain a model profile that OMI would measure for the modeled state of the atmosphere in the absence of other errors. More details about the method of model–OMI data co-location and convolution of model profiles with averaging kernel and a priori files can be found in Kumar et al. (2012).

4 Results and discussion

4.1 Model evaluation

Dust storms in northern India are characterized by large increase in aerosol optical depth (AOD) (> 50 %) and decrease in α (> 70 %) (Dey et al., 2004; Prasad and Singh, 2007). Both of these features were observed during this dust storm indicating that this was a typical pre-monsoon season dust storm (see Kumar et al., 2014) and can be considered as representative of dust storms in northern India. Kumar et al. (2014) also present a detailed evaluation of simulated aerosol optical properties, and here we present a summary. The simulated AOD, Angström exponent and single scattering albedo are compared against Aerosol Robotic Network (AERONET) measurements at seven sites in the

Table 2. Simulation designs used in this study and their purpose are shown.

Simulation index	Model configuration	Purpose
No_Dust	WRF-Chem without dust emissions but other aerosols affects photolysis rates	Serves as a base simulation
Dust_J	WRF-Chem with dust emissions and all aerosols affects photolysis rates	Comparison with No_Dust will quantify the effect of dust on photolysis rates and tropospheric chemistry
Dust_JH	Same as Dust_J but with heterogeneous chemistry considering RH dependence of γ and renoxification process included	Comparison with No_Dust will yield total impact of dust on tropospheric chemistry, and comparison with DUST_J will yield contribution of heterogeneous chemistry
Dust_JH_NoRH	Same as Dust_JH but RH dependence of γ is not included	Comparison with Dust_JH will provide information on RH dependence of heterogeneous uptake of gases by dust
Dust_JH_NoReNO _x	Same as Dust_JH but renoxification process is not included	Comparison with Dust_JH will provide contribution of renoxification process in NO _x budget.
Dust_JH_LoG	Same as Dust_JH but with lower limit of γ	Comparison with Dust_JH_HiG provides uncertainty in the effects of
Dust_JH_HiG	Same as Dust_JH but with upper limit of γ	heterogeneous chemistry due to uncertainty in γ
Dust_JH_NO ₃	Same as Dust_JH but the effect of nitrate-coated dust on uptake of O ₃ is included	Comparison with Dust_JH will provide information on changes in O ₃ uptake due to coating of dust with nitrate
Dust_JH_SO ₄	Same as Dust_JH but the effect of sulfate-coated dust on uptake of O ₃ is included	Comparison with Dust_JH will provide information on changes in O ₃ uptake due to coating of dust with sulfate
Dust_JH_NO ₃ _SO ₄	Same as Dust_JH but the effect of nitrate- and sulfate-coated dust on uptake of O ₃ is included	Comparison with Dust_JH will provide information on changes in O ₃ uptake due to coating of dust with nitrate and sulfate
Dust_JH_Sat	Same as Dust_JH but nitrate- and sulfate-coated particles are assumed to be saturated	Comparison with Dust_JH will provide upper limit on the effect of dust aging on the tropospheric chemistry

model domain. It is found that the model generally underestimates the AOD over the model domain but is able to capture the temporal variations ($r = 0.5\text{--}0.88$) seen in AERONET measurements. The good agreement between modeled and AERONET observed Angström exponent indicates that the model is able to capture dust-storm-induced variations in aerosol size. The comparison of model results with Moderate Resolution Imaging Spectroradiometer (MODIS) AOD retrievals shows that the model is also able to capture the spatial pattern of dust-storm-induced changes in the MODIS AOD as well as the spatial pattern of the dust plume. The average MODIS and WRF-Chem AOD (550 nm) values in

the high dust-laden region are estimated as 0.80 ± 0.30 and 0.68 ± 0.28 respectively.

The variations in observed and WRF-Chem-simulated (No_Dust, Dust_J, Dust_JH_NoRH and DUST_JH) daily average O₃ and NO_y at Nainital during 13–24 April 2010 are shown in Fig. 2. The modeled time series of relative humidity and dust mass concentration for particles of 0.73, 1.4 and 8.0 μm effective radii at Nainital are also shown. Note that the WRF-Chem model at the resolution ($30 \times 30 \text{ km}^2$) used here is not able to capture the rapidly varying topography around Nainital, and the altitude of the site in the model is off by about 900 m. To minimize the comparison errors induced by this spatial mismatch, we first obtain the altitude

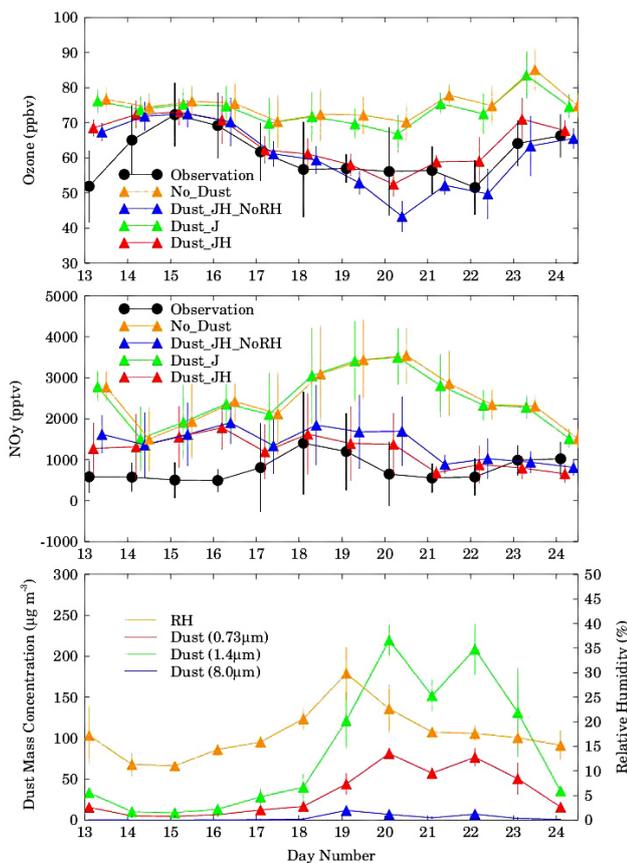


Figure 2. Variations in observed and WRF-Chem-simulated (No_Dust, Dust_JH_NoRH, Dust_J and Dust_JH) daily average O_3 (top panel) and NO_y (middle panel) at Nainital during 13–24 April 2010. WRF-Chem-simulated daily average mass concentration of dust particles of 0.73, 1.4 and $8.0\ \mu\text{m}$ effective radii and relative humidity at Nainital are also shown. The vertical bars represent standard deviation in the average values.

profile of model results at Nainital by bi-linearly interpolating model output at each model level to the location of Nainital (79.45°E , 29.36°N) and then linearly interpolate the resulting altitude profile to the height of Nainital (1958 m). Dust mass concentrations started increasing at Nainital on 17 April 2010, reached a maximum during 20–22 April and decreased thereafter. The mass concentrations of $8.0\ \mu\text{m}$ particles remain close to zero even during the dust storm because these particles do not travel far from the source regions due to their higher deposition velocities resulting in dry deposition.

The observed O_3 mixing ratios show a decrease during the dust storm period (17–22 April 2010), while observed NO_y mixing ratios show an increase during 17–18 April 2010 and a decrease thereafter. Analysis of back-air trajectories arriving at Nainital (not shown) showed that the increase in NO_y levels at Nainital during 17–18 April was associated with passage of low-altitude air masses over the Indo-Gangetic Plain region before arriving at Nainital, while

air masses on previous days (15–16 April) passed over relatively cleaner regions to the north and west at higher altitudes. The WRF-Chem model without incorporating the effects of dust aerosols (No_Dust) does not capture the observed decrease in O_3 and NO_y levels at Nainital. Including effects of dust on photolysis rates (Dust_J) induces a slight decrease (about 2 ppbv) in modeled O_3 levels, but modeled values are still significantly higher than the observations. The introduction of heterogeneous chemistry without including RH dependence of γ (Dust_JH_NoRH) in the model leads to the observed decreases in O_3 and NO_y , but compared to measurements we find too strong of decreases in O_3 levels and slightly higher NO_y levels than the observations at Nainital during the peak of dust storm, i.e., 20–22 April 2010. The inclusion of heterogeneous chemistry with RH dependence of γ (Dust_JH) leads to the best agreement between WRF-Chem-simulated and observed O_3 and NO_y values at Nainital. The WRF-Chem-simulated average O_3 values at Nainital in Dust_JH and No_Dust configurations during 17–22 April 2010 are estimated to be 58 ± 5 ppbv and 73 ± 6 ppbv respectively as compared to the average observed value of 56 ± 10 ppbv. The corresponding WRF-Chem NO_y average values at Nainital are estimated as 1189 ± 751 pptv and 2945 ± 876 pptv, respectively, as compared to the average observed value of 843 ± 887 pptv. Thus, including the effects of dust aerosols in WRF-Chem reduces the difference between average modeled and observed O_3 from 16 ± 9 to 2 ± 8 ppbv and that in NO_y from 2102 ± 1425 to 346 ± 1225 pptv respectively.

The spatial distributions of average OMI-retrieved and WRF-Chem-simulated (Dust_JH and No_Dust) tropospheric column NO_2 during 13–16 April 2010 and 17–22 April 2010 are shown in Fig. 3. The percentage differences in tropospheric column NO_2 between high and low dust emission periods are also shown for both OMI and WRF-Chem. Both the model and satellite data show similar spatial distributions with the highest values along the Indo-Gangetic Plain region during both low and high dust emission periods, but WRF-Chem generally overestimates the OMI retrievals which is consistent with previous studies over the Indian region (Kumar et al., 2012; Ghude et al., 2013). However, the comparison between Dust_JH and No_Dust configurations of WRF-Chem shows that the inclusion of effects of dust aerosols improves the model performance and reduces the model bias with respect to OMI retrievals by up to 30% especially in the Indo-Gangetic Plain region. The remaining bias in the model could be due to uncertainties in NO_x emission estimates in this region. The percentage difference plots show that the WRF-Chem model is able to capture several features of changes in OMI-retrieved spatial distribution of tropospheric column NO_2 between high and low dust emission periods. The domain-averaged OMI and WRF-Chem (Dust_JH) tropospheric column NO_2 values over the geographical region (70 – 80°E , 25 – 30°E) of maximum dust storm impact during low dust emission period are estimated

as $(2.35 \pm 1.43) \times 10^{15}$ and $(3.95 \pm 2.43) \times 10^{15}$ molecules cm^{-2} respectively, and during high dust emission period are estimated as $(2.01 \pm 1.37) \times 10^{15}$ and $(3.41 \pm 2.80) \times 10^{15}$ molecules cm^{-2} respectively. The reduction in tropospheric column NO_2 during high dust emission period in both OMI and WRF-Chem indicates that the dust storm acted as a sink for NO_2 .

4.2 Impact of dust storm on photolysis rate coefficients

The impact of the dust storm on photolysis rates is examined by comparing the daytime (07:30–17:30 IST or 02:00–12:00 UTC) NO_2 photolysis rate coefficients calculated by the WRF-Chem model with Dust_J and No_Dust configurations (Fig. 4). NO_2 photolysis rates at the surface show a strong relationship with aerosol loading and are lowest over the Indo-Gangetic Plain region, which is where the anthropogenic emissions are stronger than those over other parts of the model domain. The inclusion of dust aerosols enhances the spatial heterogeneity of NO_2 photolysis rate and decreases it by 5–25 % over the Thar Desert and western Indo-Gangetic Plain region. The photolysis rate coefficients of other trace gases such as O_3 , HNO_3 , H_2O_2 , CH_2O and N_2O_5 at the surface exhibit similar features (not shown) with decreases of the same order of magnitude. The magnitude of change in photolysis rates decreases with altitude and changes sign from negative to positive near 4 km because of the increase in actinic flux due to scattering of incoming solar radiation by dust aerosol layers underneath. The spatial structure of changes in photolysis rates at 100 hPa is similar to that at the surface with the largest increases (1–5 %) over the Thar Desert and western Indo-Gangetic Plain region.

4.3 Impact of dust storm on trace gases at the surface

To determine the impact of the dust storm on surface composition, average surface mixing ratios of O_3 , SO_2 , NO_x , HNO_3 and H_2O_2 in the DUST_JH, which is the simulation that compared best with observations at Nainital, and the No_Dust simulation are compared for the dust event time period of 17–22 April 2010 (Fig. 5). Surface O_3 shows a similar spatial distribution in both runs with the lowest values over the Arabian Sea and the highest values over the eastern Indo-Gangetic Plain region. However, the dust storm clearly leads to a reduction in O_3 mixing ratios by 3–14 ppbv (5–25 %) over the Thar Desert and western Indo-Gangetic Plain region. The spatial distribution of reductions in O_3 mixing ratios is consistent with the distribution of dust over the model domain, and the amount of reduction is comparable to those reported by previous studies in dust source regions (e.g., Dentener et al., 1996; Tang et al., 2004; Pozzoli et al., 2008; Wang et al., 2012).

Sulfur dioxide and NO_x mixing ratios are highest along the Indo-Gangetic Plain region due to higher emissions in this region (Fig. 5). Changes in SO_2 mixing ratios show a

mixed response to heterogeneous chemistry with a decrease of 0.1–0.2 ppbv (8–10 %) over the Thar Desert region and an increase of 0.2–0.5 ppbv (2–6 %) over the eastern Indo-Gangetic Plain region. This is because the heterogeneous chemistry reduces SO_2 through reaction with dust particles while increases it through reduction of OH mixing ratios ($\text{SO}_2 + \text{OH} \rightarrow \text{sulfate}$). The sign of the changes in SO_2 is thus determined by competition between these two reactions. In general, NO_x mixing ratios show a reduction of up to 0.5 ppbv (< 10 %) along the Indo-Gangetic Plain region and 0.2–0.3 ppbv (10–20 %) over the Thar Desert due to uptake of NO_2 by dust particles. The changes in NO_2 are also determined by the competition between the reactions of NO_2 with dust and OH where the former tends to decrease NO_2 while the latter tends to increase NO_2 due to reduced OH. The reduction in NO_x is in contrast with the results of Wang et al. (2012), who reported an increase in NO_x in the dust source region and attributed the increase to the renoxification process. Our study differs because Wang et al. (2012) assumed that renoxification process is active all the time, whereas this process is active only during daytime in our simulations. To quantify the contribution of renoxification process in the NO_x budget, we compared the NO_x distributions simulated by Dust_JH and Dust_JH_NoReNO_x configurations. The comparison (not shown) showed that the renoxification process does increase NO_x mixing ratios, but the magnitude of this increase (0.1–0.2 ppbv) is likely less than the reduction due to heterogeneous chemistry.

The spatial distributions of HNO_3 and H_2O_2 (Fig. 5) are largely modified by the dust storm with reduction of up to 2 ppbv (~ 99 %), because these species are highly reactive with dust particles (Table 1). The large reduction in H_2O_2 mixing ratio estimated here is also in contrast with the results of Wang et al. (2012), who estimated an increase in H_2O_2 . Wang et al. (2012) assumed a 100 % yield for the conversion of HO_2 into H_2O_2 through heterogeneous uptake, while we assume a yield of 10 % following recent work by Bedjanian et al. (2013b). Mixing ratios of NO_3 , N_2O_5 and CH_3COOH also show large reductions by Dust_JH with the highest decrease reaching up to 0.1 ppbv (~ 98 %), 0.46 ppbv (~ 99 %) and 0.45 ppbv (~ 96 %) respectively. The uptake of HO_2 by dust particles leads to a maximum reduction of about 3.5 pptv (~ 40 %) over the Thar Desert region, which is much less than those reported previously (e.g., Bian and Zender, 2003; Wang et al., 2012) and is attributed to a lower γ value used in our study (< 0.064 vs. 0.1). The maximum reduction in OH, CH_2O and CH_3OH is estimated to be about 40, 21 and 5 %, respectively.

The above discussion includes both the role of heterogeneous chemistry in changing the distribution of trace gases and dust-modified photolysis rate coefficients. The individual contributions of heterogeneous chemistry and perturbation in photolysis rate coefficients to the total difference in distributions of trace gases at the surface are estimated by comparing differences between simulations (Fig. 6). More

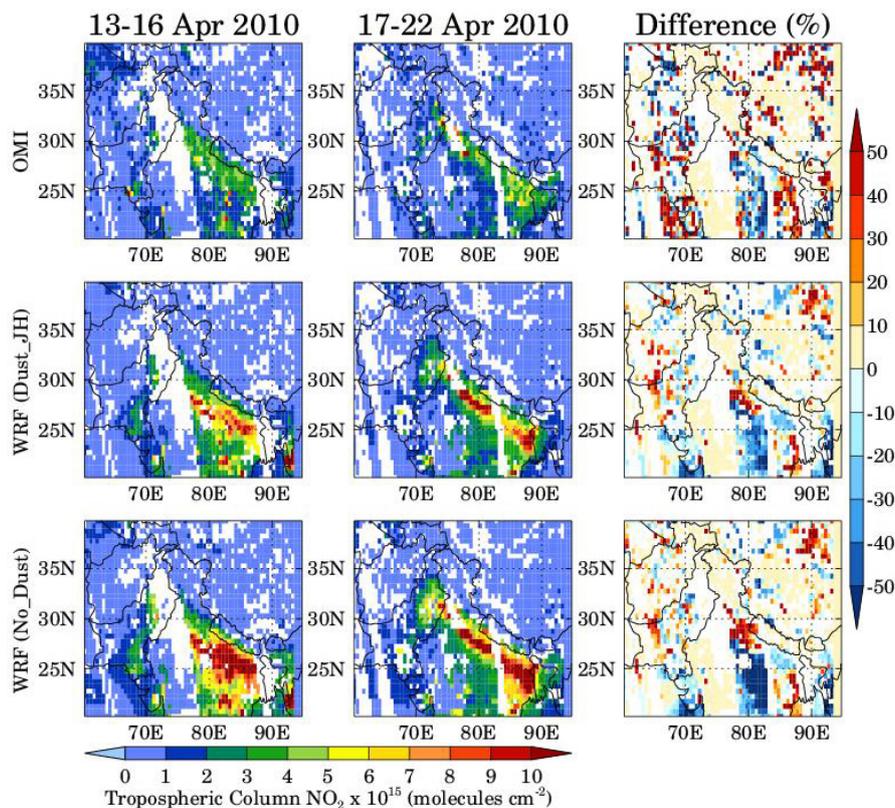


Figure 3. Spatial distributions of OMI-retrieved and WRF-Chem-simulated (Dust_JH and No_Dust) tropospheric column NO_2 during the low (13–16 April 2010) and high (17–22 April 2010) dust emission periods. WRF-Chem profiles are convolved with OMI averaging kernels before comparison. The percentage differences between low and high dust emission periods are also shown.

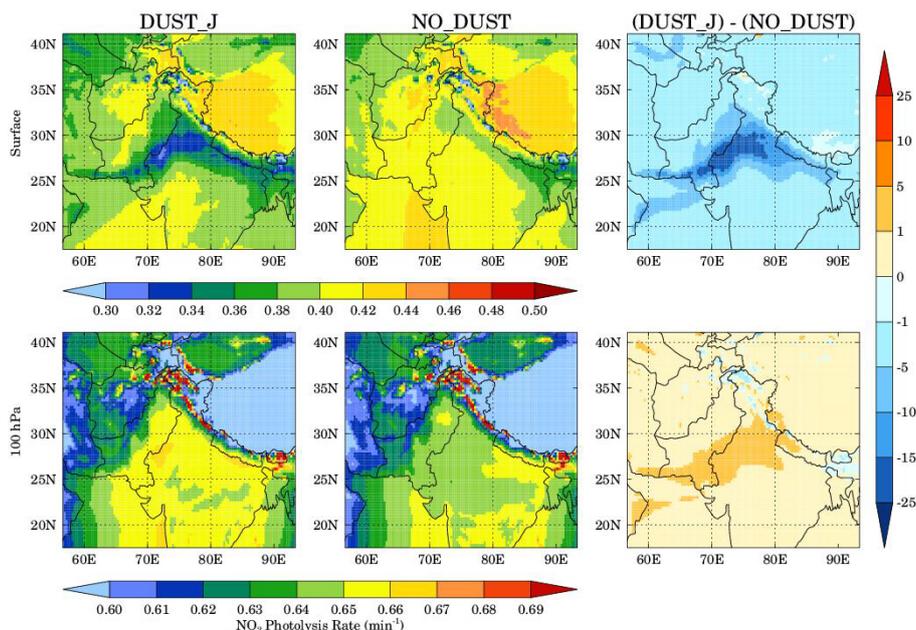


Figure 4. Spatial distributions of the WRF-Chem-simulated daytime (02:00–12:00 UTC) NO_2 photolysis rate with (Dust_J) and without dust (No_Dust) at the surface (top panel) and 100 hPa (bottom panel) during 17–22 April 2010 are shown. Percentage differences between Dust_J and No_Dust cases are also shown.

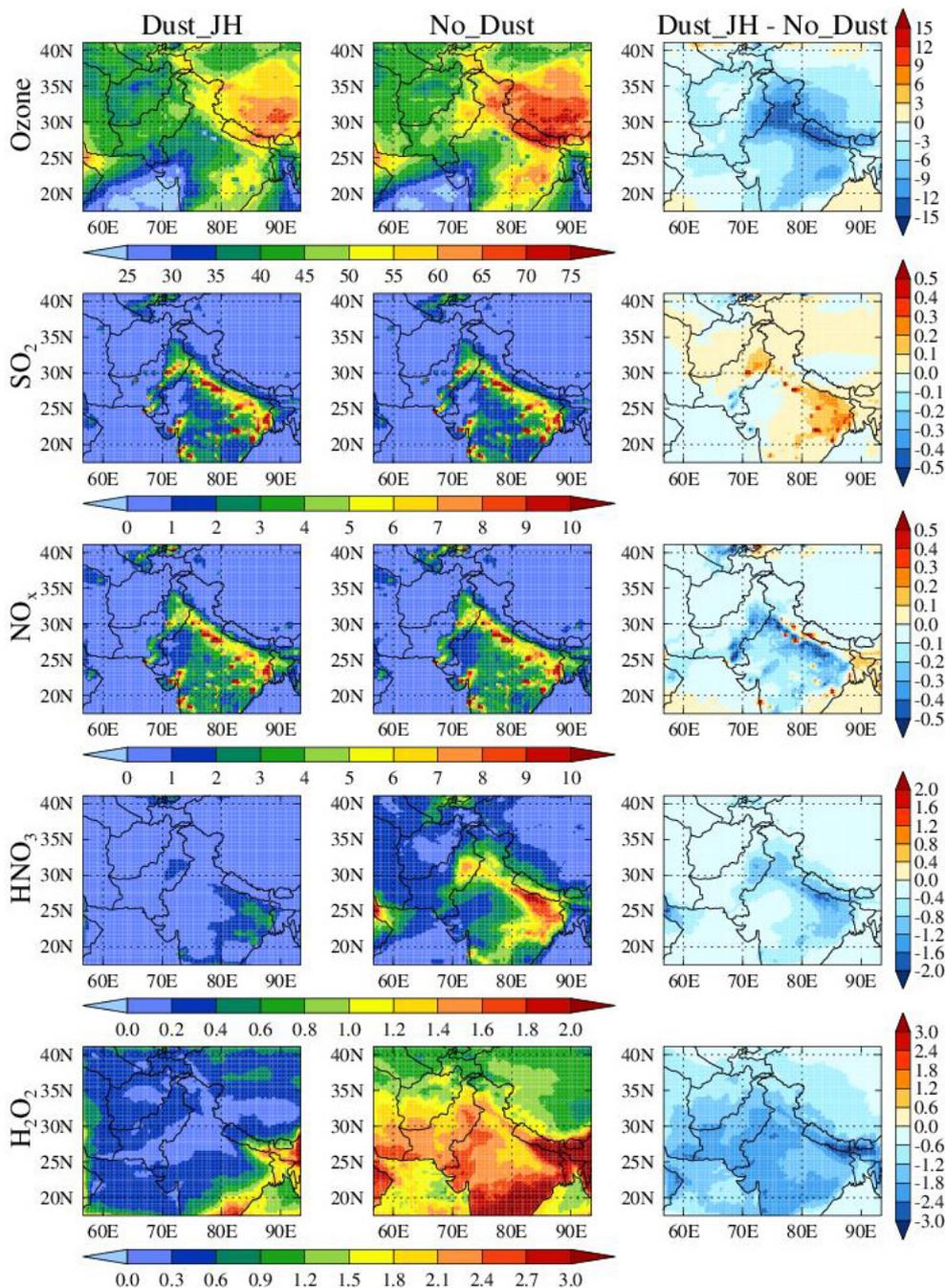


Figure 5. Spatial distributions of average surface O₃, SO₂, NO_x, HNO₃ and H₂O₂ mixing ratios simulated by WRF-Chem with Dust_JH and No_Dust configuration during 17–22 April 2010. The absolute difference between the two configurations is also shown. All values are in ppbv.

than 80% of the changes in distribution of these trace gases are explained by the heterogeneous chemistry. The changes in surface O₃ induced by dust-modified photolysis rate coefficients are within ± 1 –3 ppbv and are driven by the complex response of O₃ to decreases in photolysis rate coefficients. A decrease in the photolysis rate coefficients leads to slower

photochemical processing in general and thus decreased O₃ production.

Mixing ratios of surface SO₂ and NO_x show an increase of up to 0.1 ppbv due to reduction in photolysis rate coefficients. The increase in NO_x and SO₂ is associated with a decrease in OH mixing ratios as reaction with OH is the main loss process for both NO_x and SO₂. The decrease in O₃

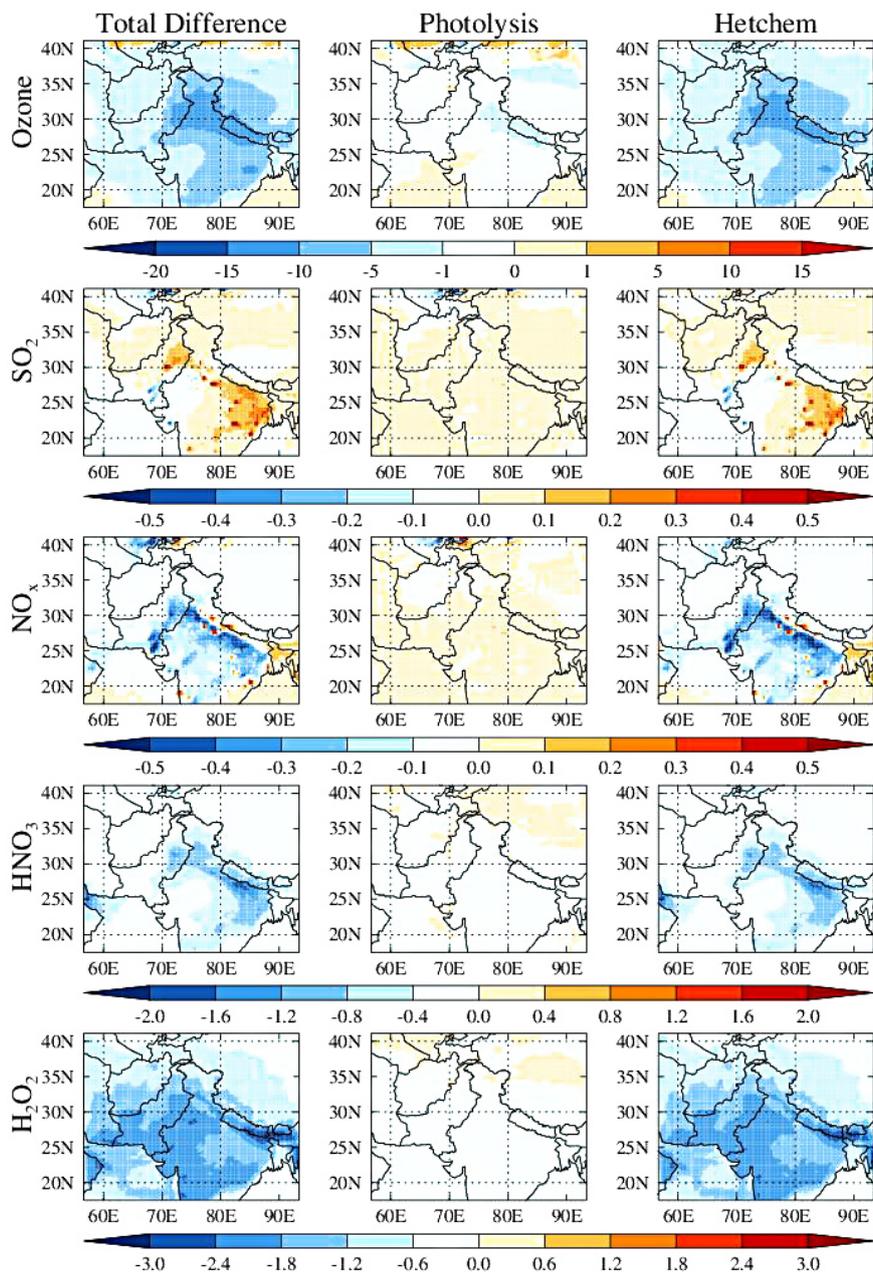


Figure 6. Spatial distributions of absolute difference in average O₃, SO₂, NO_x, HNO₃ and H₂O₂ between Dust_JH and No_Dust (left panel), Dust_J and No_Dust representing the contribution of photolysis to total difference (middle panel), and Dust_JH and Dust_J representing the contribution of heterogeneous chemistry to total differences (right panel). All values are in ppbv and for the surface layer of the model.

photolysis rate coefficient leads to a decrease of up to 30 % in OH ($\text{O}_3 + h\nu \rightarrow \text{O}^1\text{D} + \text{O}_2$, producing OH through reaction of O¹D with water vapor) mixing ratios. Mixing ratios of surface HNO₃ and H₂O₂ also show a small decrease of up to 0.4 and 0.6 ppbv in the Dust_J configuration relative to the No_Dust configuration. HNO₃ is produced mainly by the reaction of OH with NO₂, but the rate of change of HNO₃ is dominated by changes in OH as the reduction in OH (up to 30 %) is larger than that in NO₂ (up to 5 %). The reduction

in H₂O₂ is also associated with reduction in OH and HO₂ mixing ratios. Since OH is the major oxidizing agent in the troposphere, a decrease in OH also leads to a decrease in the oxidizing capacity of the atmosphere and a consequent increase of up to 5–10 % in several trace gases such as CO, alkanes and alkenes at the surface.

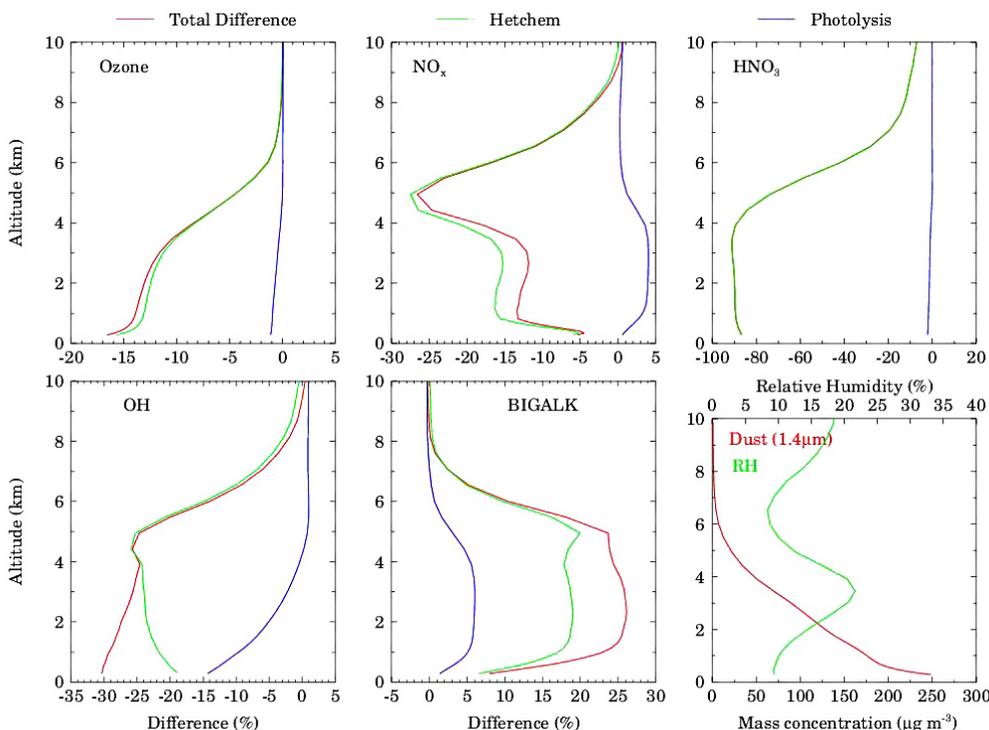


Figure 7. Vertical profiles of percentage total difference (red lines) in average O_3 , NO_x , HNO_3 , OH and BIGALK between Dust_JH and No_Dust configuration over the geographical region ($70\text{--}80^\circ\text{ E}$, $25\text{--}30^\circ\text{ N}$) of maximum dust influence. The contributions of heterogeneous chemistry (green lines) and photolysis (blue lines) to the total percentage difference are also shown. The vertical profiles of dust mass mixing ratios for particles of $1.4\ \mu\text{m}$ effective radius and relative humidity are also shown to help interpretation.

4.4 Impact of dust storm on vertical distribution of trace gases

In addition to changes in surface mixing ratios, vertical distributions of trace gases are modified by the dust storm via heterogeneous chemistry and dust-modified photolysis rates. By averaging over the region of maximum dust influence ($70\text{--}80^\circ\text{ E}$, $25\text{--}30^\circ\text{ N}$) for each simulation, the percentage differences between Dust_JH and No_Dust giving the total difference, Dust_J and No_Dust giving the contribution of modified photolysis rates, and Dust_JH and Dust_J giving the contribution from heterogeneous chemistry are found. Profiles of O_3 , NO_x , HNO_3 , OH and BIGALK are shown in Fig. 7. BIGALK represents alkanes with four or more carbon atoms in the MOZCART chemical mechanism and is shown to illustrate the dust-storm-induced changes in volatile organic compounds.

The changes in all the trace gases due to heterogeneous chemistry are much larger than those due to perturbations in photolysis rate coefficients, and are significant below 8 km, which is consistent with the vertical distribution of dust particles (Fig. 7). All gases except BIGALK show a net reduction because of the larger changes induced by the heterogeneous chemistry. The reduction in OH due to decrease in O_3 photolysis rate coefficient leads to a small increase of 1–4 %

in NO_x . The reaction with OH is the only loss process for BIGALK, and therefore BIGALK shows an increase due to both heterogeneous chemistry and perturbation in photolysis rates as both of these processes lead to a decrease in OH. The highest net decreases in O_3 , NO_x , HNO_3 and OH are estimated as ~ 16 , ~ 26 , ~ 91 and ~ 30 % respectively while the highest net increase in BIGALK is estimated as ~ 26 %. The vertical distributions of changes in NO_3 , N_2O_5 , H_2O_2 and CH_3COOH are similar to those in HNO_3 , and the highest net reduction reaches 80–90 %.

4.5 Importance of RH dependence of reactive uptake coefficients (γ)

The uptake of trace gases by dust aerosols also depends upon the relative humidity as reactive uptake coefficients have a large variation with RH. The effect of relative humidity on the rate constants k_{O_3} and k_{HNO_3} at the surface over the model domain during 17–22 April 2010 is illustrated in Fig. 8, where model runs with and without RH dependence of γ are compared. The spatial distributions of average mass concentration for dust particles of $1.4\ \mu\text{m}$ effective radius and relative humidity are also shown to help the interpretation. RH is less than 20 % over most of the Indian region and is 70–90 % over the oceanic regions. As expected, the spatial distributions of both k_{O_3} and k_{HNO_3} in both configurations are

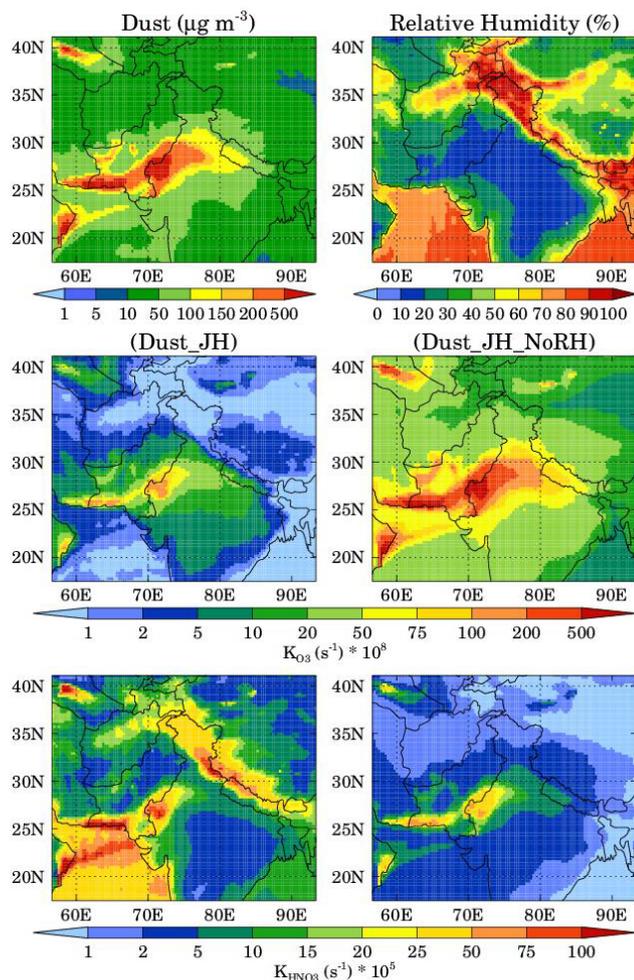


Figure 8. Spatial distributions of average mass concentration for dust particles of $1.4\mu\text{m}$ effective radius, relative humidity and pseudo-first-order rate coefficients for heterogeneous uptake of O_3 (k_{O_3}) and HNO_3 (k_{HNO_3}) by dust particles in Dust_JH and Dust_JH_NoRH configurations during 17–22 April 2010.

nearly identical to the distribution of dust mass concentrations with the highest values in the dust source regions. The rate constant decreases for O_3 while it increases for HNO_3 by 1–2 orders of magnitude after inclusion of RH dependence of γ . The rate coefficients for other gases show a similar spatial distribution.

The effect of RH-induced changes in the rate constants on heterogeneous uptake of O_3 , SO_2 , HNO_3 , H_2O_2 , OH and HO_2 is illustrated by showing the relative percentage differences in the distribution of these gases in the lowest model layer between the model runs with and without RH dependence of γ (Fig. 9). Surface O_3 and HO_2 show an increase when RH effects on γ are included because γ for these gases decreases with RH. In contrast, HNO_3 and H_2O_2 uptake coefficients increase when RH effects are accounted for, resulting in a decrease in HNO_3 and H_2O_2 mixing ra-

tios. The changes in O_3 , HO_2 and H_2O_2 reach up to 20, 25 and 50 %, respectively, over the Thar Desert and the western Indo-Gangetic Plain region. The percentage changes in HNO_3 reach up to 100 % and are higher over the oceanic regions ($\text{RH} > 45\%$) than the inland regions because of the order of magnitude increase of γ for HNO_3 when RH increases above 40 %. Surface SO_2 and OH show a mixed response with increase over some parts of the model domain and decrease over the others. The changes in SO_2 and OH are within $\pm 10\%$. The changes in SO_2 are determined by the competition between the reactions of SO_2 with dust particles and OH. The heterogeneous reaction of SO_2 tries to reduce SO_2 due to increase in γ with RH while that with OH would decrease (increase) SO_2 if OH increases (decreases). The magnitudes of these RH-induced changes in trace gases are comparable to those induced by heterogeneous chemistry (reported in Sect. 4.3). This suggests that consideration of RH-dependent γ values in heterogeneous chemistry calculations is as important as is the accurate simulation of dust mass concentrations.

4.6 Effect of uncertainty in reactive uptake coefficient (γ)

The effect of uncertainty in reactive uptake coefficient (γ) on heterogeneous chemistry induced changes in the tropospheric chemistry is illustrated in Fig. 10, where model runs with lower (Dust_JH_LoG) and upper (Dust_JH_HiG) bounds of γ are compared in the lowest model layer. It is clear that uncertainty in γ can lead to significant uncertainty in heterogeneous chemistry induced changes in all gases. Surface O_3 mixing ratios have uncertainties of up to 8–11 ppbv (25–30 %), while other gases have uncertainties of up to 1–1.5 ppbv (30–100 %), except SO_2 and NO_2 , which show uncertainties of up to 3 ppbv (30–70 %) at some locations. A comparison of these uncertainties with the total change induced by dust storm in these gases (Fig. 5 and Sect. 4.3) reveals that uncertainties in O_3 , HNO_3 , NO_3 , N_2O_5 , OH, HO_2 , H_2O_2 , and CH_3COOH due to the uptake values have magnitudes similar to the total changes induced by dust aerosols, while those in NO_2 , SO_2 and CH_3OH are even greater than the total changes induced by dust storm. These results highlight the importance and necessity of accurate measurements of reactive uptake coefficients.

4.7 Impact of dust aging

The aging of dust particles through heterogeneous uptake of gases can modify the dust reactivity towards trace gases. We have made an attempt to simulate changes in the dust reactivity and uptake of trace gases due to nitrate and sulfate coating, and the results are presented in this section. The spatial distributions of WRF-Chem-simulated average mass concentration of fresh and aged dust particles of $1.4\mu\text{m}$ effective radius at the surface in Dust_JH_NO₃, Dust_JH_SO₄

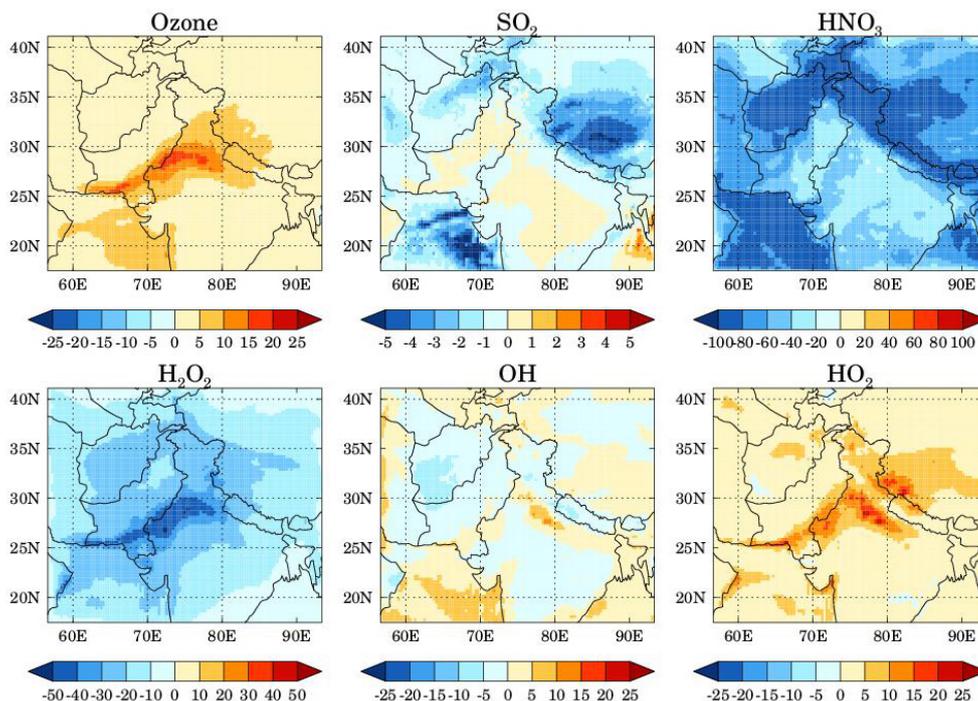


Figure 9. Spatial distributions of relative percentage differences in O_3 , SO_2 , HNO_3 , H_2O_2 , OH and HO_2 mixing ratios at the surface between Dust_JH and Dust_JH_NoRH configurations during 17–22 April 2010.

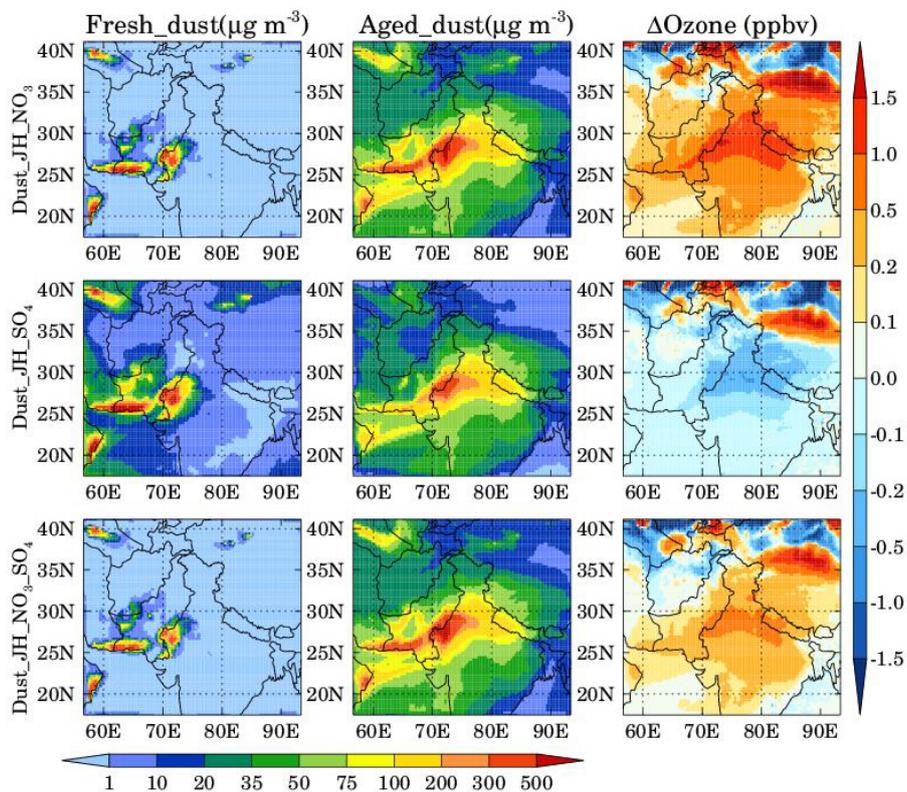


Figure 10. Spatial distributions of absolute differences in O_3 , HNO_3 , NO_2 , SO_2 , CH_3OH and CH_2O surface mixing ratios between Dust_JH_LoG and Dust_JH_HiG configurations during 17–22 April 2010. All values are in ppbv.

and Dust_JH_NO₃_SO₄ configurations during 17–22 April 2010 are shown in Fig. 11. In general, fresh dust particles are seen mostly in and near the source regions because all dust particles are emitted as fresh particles and have much smaller concentrations than the aged dust particles particularly outside the dust source regions. Among all the gases providing nitrate coating on dust particles, the HNO₃ uptake makes the highest contribution to the concentration of aged particles due to its higher mixing ratios and strong increase in its reactivity towards the dust surface with relative humidity. For example, if there are 10 fresh dust particles cm⁻³ of effective radius 0.73 μm, then the uptake of HNO₃, NO₂, NO₃ and N₂O₅ leads to about 2.5 aged dust particles cm⁻³ in one time step (180 s) at RH = 0, and HNO₃ and NO₃ each provide about 1 dust particle. However, the number concentration of aged dust particles formed due to HNO₃ uptake increases to about 7 particles cm⁻³ per time step as RH increases to 20–40%. The uptake of SO₂ also increases with increase in relative humidity, but the SO₂ contribution to the aged particles is much less than that of HNO₃ due to lower SO₂γ values. The decrease in O₃γ values for nitrate-coated dust particles leads to an enhancement of 1–2 ppbv in surface O₃ over the Thar Desert and western Indo-Gangetic Plain region relative to uncoated dust particles (Fig. 11) while increase in O₃ γ values for sulfate-coated dust particles leads to a reduction of 0.5–1 ppbv over these regions. Since nitrate coating leads to a larger fraction of aged particles than the sulfate coating, surface O₃ mixing ratios show an overall enhancement of up to 1 ppbv over the regions of high dust loadings when both sulfate- and nitrate-coated dust particles are allowed to react with O₃ in the model.

The presence of a nitrate or sulfate monolayer on the dust particle might saturate the dust particles and deactivate them for further catalytic uptake of other gases. In order to examine the effect of such a coating on the uptake of O₃, H₂O₂, HCHO and CH₃COOH, lower and upper bounds of heterogeneous chemistry reaction rates induced changes in the surface mixing ratios of these gases are calculated (Fig. 12). Lower and upper bounds for each gas are calculated by subtracting their average values in Dust_J configuration from those in Dust_JH and Dust_JH_Sat configurations respectively. Absolute mixing ratios of these gases in Dust_J configuration, in which dust aerosols affected photolysis rates only, are also shown to provide an idea of the modification in base levels of these gases due to heterogeneous chemistry. As expected, the saturation of dust particles decreases the magnitude of reduction caused by heterogeneous chemistry for all the gases by 5–50%. The amount of maximum reduction in surface O₃ changed from 20–25% to 15–20% when saturation effects are accounted for. The saturation of dust particles has a larger impact on the distribution of H₂O₂ and CH₃COOH as maximum reduction in both of these gases decreased to 20–40% (as compared to 70–90% for unsaturated dust particles) over the Thar Desert and dust source regions. Both of these gases show a small increase of 0.1–0.2 ppbv (less

than 10%) outside the source regions for the case of saturated dust particles. The maximum reduction in HO₂ and OH (not shown) also changes from 20–40% in Dust_JH to 5–20% in Dust_JH_Sat. Some changes can also be discerned in the distribution of HCHO and CH₃OH (not shown), but they are within ±5%.

5 Summary

The effects of a typical pre-monsoon season dust storm on tropospheric chemistry are analyzed for a case study in northern India, using the Weather Research and Forecasting model coupled with Chemistry (WRF-Chem), which is further developed to enhance its ability to simulate tropospheric chemistry in the presence of dust particles. The changes made to the model are specific to the MOZCART setup. Two major updates are included in this study: firstly, the F-TUV photolysis scheme of the model is updated to include the effect of dust aerosols on photolysis rates and to achieve consistency between the methods through which aerosols affect the meteorology and photolysis rates in the model. Secondly, a new scheme consisting of 12 heterogeneous reactions is included to simulate heterogeneous chemistry on the surface of dust particles. The relative humidity dependence of uptake coefficients, which was ignored in most previous studies, is used for six of the heterogeneous reactions.

The extended configuration of WRF-Chem is applied to a typical pre-monsoon season dust storm that occurred in northern India during 17–22 April 2010. The model reproduced the spatial and temporal distribution of dust plumes and aerosol optical properties (Kumar et al., 2014). The simulations are evaluated against surface O₃ and NO_y observations at a high-altitude (1958 m) measurement station in the Himalayan region (Nainital), and the model is found to capture the observed decrease in O₃ and NO_y during the dust storm only after the inclusion of the effects of dust on photolysis rates and heterogeneous chemistry. Average observed and modeled O₃ values at Nainital during 17–22 April 2010 are estimated to be 56 ± 10 and 58 ± 5 ppbv respectively, and the corresponding NO_y values are estimated to be 843 ± 887 and 1189 ± 751 pptv respectively. The extended configuration of the model also reduced biases in tropospheric column NO₂ by up to 30% compared to OMI retrievals and captured the general features of the dust-storm-induced changes in the spatial distribution of OMI-retrieved tropospheric column NO₂.

Several sensitivity simulations are conducted to investigate the contribution of different processes on mixing ratios of several key trace gases including ozone, nitrogen oxides, hydrogen oxides, methanol, acetic acid and formaldehyde. The dust storm leads to a decrease of 5–25% in photolysis rate coefficients of O₃, NO₂ and other trace gases at the surface and an increase of 1–5% above 4 km. It is found to have a significant impact on the regional tropospheric chemistry: a

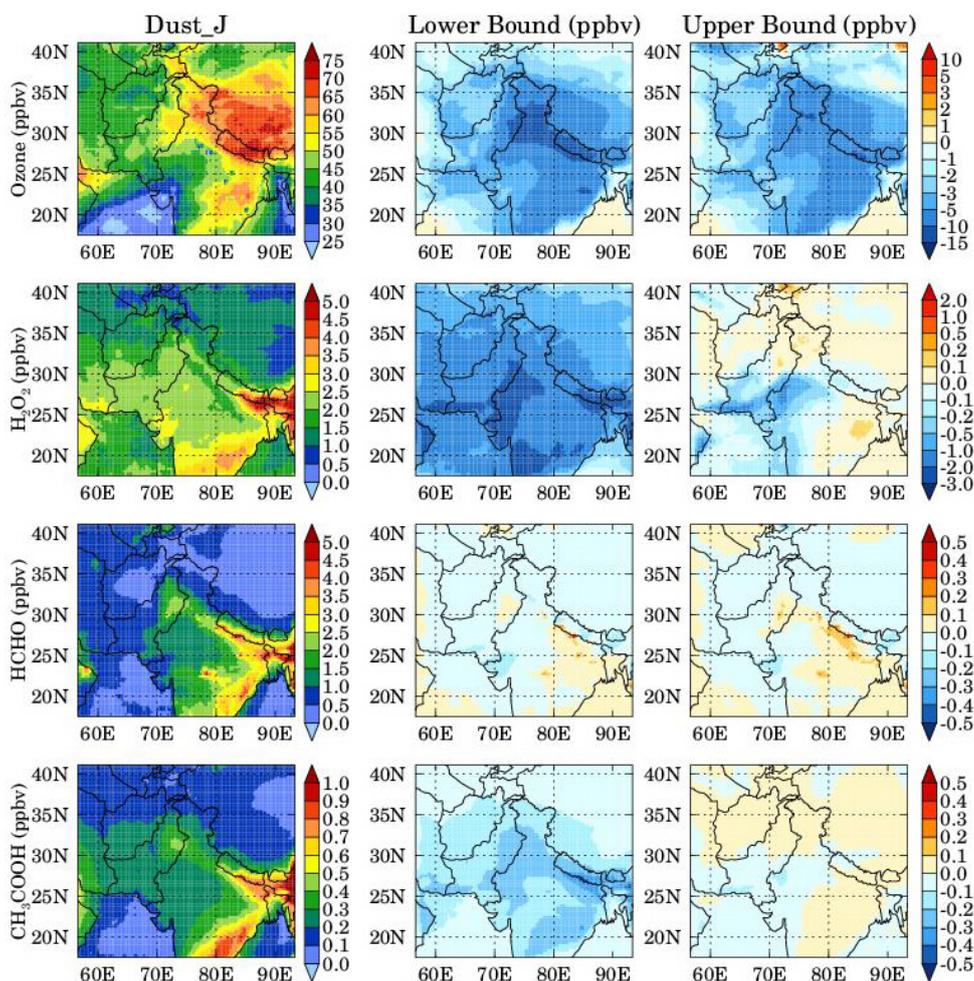


Figure 11. Spatial distributions of WRF-Chem-simulated average mass concentration of fresh and aged dust particles of $1.4\ \mu\text{m}$ effective radius at the surface in Dust_JH_NO₃, Dust_JH_SO₄ and Dust_JH_NO₃_SO₄ configurations during 17–22 April 2010. ΔO_3 is calculated by subtracting the average surface O₃ in Dust_JH configuration from the corresponding average values in Dust_JH_NO₃, Dust_JH_SO₄ and Dust_JH_NO₃_SO₄ configurations respectively.

decrease of 5–99 % is estimated in the mixing ratios of a variety of trace gases including O₃, nitrogen oxides, hydrogen oxides, sulfur dioxide, methanol, acetic acid and formaldehyde at the surface. Analysis of the vertical distributions of these trace gases shows that dust-storm-induced changes are significant up to an altitude 8 km and are estimated as 80–90 % (5–10 times) for highly reactive gases such as HNO₃, NO₃, N₂O₅, H₂O₂ and CH₃COOH. It is found that the majority of these changes are induced by the heterogeneous chemistry, while the contribution of dust-modified photolysis rates generally remained less than 10 %. An increase of up to 30 % in volatile organic compounds is estimated due to decreases in OH concentrations.

The RH dependence of γ is found to play a large potential role in heterogeneous chemistry. Sensitivity studies showed that the exclusion of the RH dependence can introduce a difference of 1–2 orders of magnitude in heterogeneous reac-

tions rate constants, 20–25 % changes in O₃ and HO₂, and up to 50 and 100 % changes in H₂O₂ and HNO₃, respectively. These effects are comparable to heterogeneous chemistry induced changes in these gases. The effect of uncertainties in the uptake coefficients (γ) on the distribution of trace gases is examined using sensitivity simulations with lower and upper bounds of γ . Differences of trace gas mixing ratios caused by uncertainties in the uptake coefficients are found to be the same magnitude or greater as the differences in mixing ratios induced by dust storm effects on tropospheric chemistry. We also tested the importance of atmospheric aging of dust particles in the context of heterogeneous chemistry. Model experiments based on laboratory studies of changes in dust reactivity due to atmospheric aging showed that coating of dust with nitrate and sulfate may lead to changes of up to 1 ppbv in surface O₃ simulations. A hypothetical simulation is also conducted by saturating the nitrate- and sulfate-coated

dust particles for uptake of O_3 , HO_2 , OH , H_2O_2 , CH_3COOH , CH_3OH and $HCHO$. The saturation of dust particles is found to have a larger impact on the distributions of H_2O_2 and CH_3COOH but a relatively small impact on other gases.

This study clearly shows that the pre-monsoon season dust storm can potentially affect the regional tropospheric chemistry in northern India. However, the implications of the heterogeneous uptake of trace gases on aerosol size distributions and their feedbacks on the radiation budget and cloudiness are not examined here. Dust particles coated with nitrate/sulfate may interact differently with radiation as compared to uncoated dust particles and can increase or decrease cloudiness depending upon their size distribution. Both of these processes will have important implications for the direct aerosol radiative forcing and the elevated heat pump (EHP) hypothesis (Lau et al., 2006), which proposed that the absorption of solar radiation by dust and black carbon aerosols along the southern slopes of Himalayas modulates the meridional temperature gradient and leads to an early onset of Indian summer monsoon. The heterogeneous chemistry scheme implemented in the MOZCART chemical mechanism here can be easily extended to a more detailed aerosol module (e.g., MOSAIC) of WRF-Chem, which would allow studies on the implications of heterogeneous chemistry for aerosols and their interaction with radiation, clouds, and the Asian monsoon, including the role of aerosol aging on aerosol and trace gas distributions.

Nevertheless, this study demonstrates that the effects of dust aerosols through heterogeneous chemistry and perturbation in photolysis rates should be included in atmospheric chemistry transport models, especially for simulating air quality in northern India. Although this study analyzed a typical dust storm in northern India, but more such studies should be conducted in future to lend further confidence in these results. At the same time, it is also imperative to improve the accuracy and precision of the reactive uptake coefficients, their dependence on relative humidity and atmospheric processing of dust particles. In addition, extensive efforts must be made to conduct co-located measurements of O_3 and related gases, along with physical and chemical properties of dust aerosols in northern India, especially during the dust storm season, to gain insights into the effects of dust aerosols on tropospheric chemistry and provide more data for model evaluation.

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